

Assessment of two sludge stabilization methods in a wastewater treatment plant in Sotaquirá, Colombia

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Abstract

Wastewater sludge is a by-product of wastewater treatment; it is often employed in agricultural processes following appropriate microbiological stabilization. In this study, we evaluated two methods of sludge stabilization in the wastewater treatment plant of Sotaquirá, Colombia. The two assessed stabilization methods were dewatering on drying beds and alkaline stabilization. The stabilization through dewatering on drying bed was carried out for five months. Alkaline stabilization was conducted for 96 hours with three concentrations of calcium oxide, 7 %, 9 %, and 13 % w/w. Humidity, pH, organic carbon, phosphorus, organic nitrogen, fecal coliforms, somatic phages, total helminth eggs, and *Salmonella* sp. were monitored monthly for the sludge under dewatering treatment and at 0, 12, 24 hours for sludge under alkaline stabilization treatment. Both treatments succeeded in eliminating *Salmonella* sp. Helminth eggs were eliminated by alkaline stabilization, whereas it was reduced to one or zero helminth eggs with the dewatering treatment. Somatic phages were eliminated with alkaline stabilization but were only reduced to 3.52 log CFU/g with the dewatering method. Dewatering on drying beds produced biosolids that can be used for soil restoration. Whereas alkaline stabilization produced biosolids that can be used for agricultural purposes. Alkaline stabilization with 9 % and 13 % calcium oxide ostensibly reduced nitrogen and phosphorus contents in the sludge, whereas 7 % calcium affected less the phosphorus concentration of the sludge. These results indicate that sludge dewatering on drying beds is an effective sludge sanitation protocol to be implemented in small wastewater treatment plants, such that in Sotaquirá, Colombia.

Keywords: Drying beds; alkaline stabilization; wastewater sludge; biosolids; pathogens; organic carbon.

Introduction

Wastewater sludge is a by-product of wastewater treatment. Under ideal circumstances, this sludge is treated and disposed or applied on the soil for its development. Currently, the rapid and massive accumulation and inappropriately treated wastewater sludge is a matter of concern because of its negative effects on the environment and public health [1-3]. To minimize

these negative impacts, several countries have enacted policies that require stabilization and removal of sewage sludge prior to its soil application [4-6]. In this regard, Colombian regulations deem sewage sludge stabilization as a way to reduce pathogens and further classify its biosolids, based on their physicochemical and microbiological properties, into categories A and B (see **Table 1** and Table 3). Wastewater sludge type A biosolids can be used for agricultural purposes without restrictions. Whereas, sewage sludge type B biosolids can only be used for soil restoration [4].

In Colombia, 97 % of the sludge produced by wastewater treatment plants (WWTP) is generated in the three major urban areas, Bogotá, Cali, and Medellín [7]. The remaining 3 % is produced by several small municipal plants across the country. In these municipal WWTPs knowledge about how to appropriately manage growing sewage sludge is missing. Usually, these WWTPs dispose the sludge on nearby fields or give it away to local farmers without any treatment, oblivious of the negative impact of their practices on the environment and the public health.

The wastewater treatment plant in the locality of Sotaquirá, (Department of Boyacá) Colombia, produces 109 tons of total solids per year. The sewage sludge generated by this plant is disposed on sludge drying beds for a short period of time, usually less than two months, due to its limited storage capacity. Then, it is given away to local farmers without any further treatment. This is seen as a win-win situation for the plant managers and the farmers because the plant gets rid of the by-product and the farmers use it to increase the productivity of their crops. However, they do not realize the negative health and environmental consequences that this practice may cause. This sewage sludge exceeds the pathogenic microorganism concentrations permitted for agricultural purposes or soil restoration [4]. This problem exists even though there are several stabilization techniques that have been proven to successfully reduce the pathogenic bacteria concentration to the required levels by local regulations.

Some of the stabilization techniques available include sludge anaerobic digestion [8], drying of sewage sludge on radiant floors using solar energy and biogas [9], sludge dewatering on drying beds [10-11], composting [12], and alkaline stabilization [13-16]. In this paper, we present the results of the evaluation of two of these stabilization methods on the sewage sludge generated by the WWTP of Sotaquirá: sludge dewatering on drying beds and alkaline stabilization. These methods were chosen because of their operational simplicity, low consumption of energy, and a low level of investment. In particular, we carried out physicochemical and microbiological analyses to evaluate whether or not these methods produce biosolids suitable for agricultural use or soil restoration.

Table 1. Sewage sludge biosolids classification according to Colombian regulations, based on microbiological variables.

| Variable | Category A | Category B |
|---------------------------------|------------|------------|
| Fecal coliforms (U/g) | < 3.00 | < 6.30 |
| Somatic phages (log FU/g) | < 4.70 | - |
| Viable helminth eggs (HH/4g) | < 1.00 | < 10.00 |
| <i>Salmonella</i> sp. (CFU/25g) | 0 | < 1.000 |

Materials and methods

Description of the area of study

The wastewater treatment plant is located in Sotaquirá, in the department of Boyacá, Colombia (5°45' 4" N, 73°14' 50.9" W) at 2 860 meters above sea level. The average annual temperature of this locality is 14 °C with an average annual relative humidity of 80 %, an average annual precipitation of 1 260 mm, and an average wind speed of 21 Km/h. The main economic activity of this town is dairy production and all its discharges flow into the sewage system. The wastewater treatment plant treats an average flow of 1.72 L/s.

Collection of samples

The wastewater sludge accumulated in an Imhoff tank is uniformly distributed on one of the two sludge drying beds available. The sludge drying bed used in this study has an effective area of 5.5 m². A transparent plastic sheet was used to block rainfall without blocking solar radiation. The drying process consisted of sun exposure for five months. During this period, dewatering sludge was sampled monthly from the surface of the bed. Each sample was taken by digging small holes from several points until collecting approximately 1 Kg. The monthly samples were collected and stored in a hermetic storage bag at 4 °C for 48 h before analysis.

Laboratory analyses

Physicochemical and microbiological analyses were carried out on the collected samples. Prior to analysis, the samples were pulverized and sieved through a 1.18 mm stainless steel mesh (Endecotts, United Kingdom). As for the microbiological analysis, fecal coliforms, helminth eggs, *Salmonella* sp., and somatic phages were assessed. Whereas in the physicochemical analysis, heavy metals, total organic carbon (TOC), total phosphorus, organic nitrogen, pH, humidity, conductivity, and organic matter were quantified. Heavy metal analyses were carried out only with the sample collected at the beginning of the study. Most of the other analyses were done in triplicate every month for a five-month period. The methods used for each of the analyses are listed in **Table 2**.

A sample of 100 g of WWS was collected after two months of natural dewatering in drying sludge beds. This sample was pulverized, weighed, and thoroughly mixed with reagent grade calcium oxide (CaO) in the following three concentrations 7 %, 9 %, and 13 %, on a dry weight basis (w/w). The three sample mixtures were analyzed at 0, 12, 24, and 96 hours for total organic carbon concentration, organic matter load, total phosphorus concentration, pH, and conductivity. Total coliforms, total helminth eggs, *Salmonella* sp., and somatic phages were also assessed. All data from physical-chemical variable quantifications and microbiological test results were handled and analyzed using R [17].

Results and discussion

The concentration of heavy metals in the samples of wastewater sludge from Sotaquira's WWTP are shown in **Table 3**. The obtained concentrations of these metals are below permitted values for agricultural use without restrictions, thus this sludge can be safely assigned to category A. These results may be related to the absence of industrial and mining activity in the area of the municipality. In comparison, sludge in big cities with high industrial development presents higher values of heavy metals. For instance, readings of the WWTP of Medellin have revealed a copper in a concentration of 894.6 mg/Kg, cadmium 11.17 mg/Kg, lead 94.6 mg/Kg, and nickel 398.2 mg/Kg [18].

Physicochemical and microbiological features of the sludge on drying bed treatment

The sludge discharged onto the drying bed had a pH value of 5.5. This pH value could be explained by fermentation of organic matter by facultative anaerobic bacteria. After one month on the drying bed, the pH values increased and remained around 6.84 (± 0.20) for the rest of the drying period. These results

Table 2. Methods used for the physicochemical and microbiological analyses.

| Metal | Result (mg/kg) | Category A (mg/kg) threshold | Category B (mg/kg) threshold |
|-----------------|----------------|------------------------------|------------------------------|
| | 1.00 | 1.00 | 1.00 |
| Arsenic (As) | < 1.8 | 20 | 40 |
| Cadmium (Cd) | < 1.8 | 8 | 40 |
| Copper (Cu) | < 18 | 1000 | 1750 |
| Chromium (Cr) | < 18 | 1000 | 1500 |
| Mercury (Hg) | < LQ | 10 | 20 |
| Molybdenum (Mo) | < LQ | 18 | 75 |
| Nickel (Ni) | < 18 | 80 | 420 |
| Lead (Pb) | < 18 | 300 | 400 |
| Selenium (Se) | < 1.8 | 36 | 100 |
| Zinc (Zn) | 82 | 2000 | 2800 |

differ from the results obtained by Cota-Espericueta and Ponce-Corral using a solar dryer [19]. They observed an initial pH level of 6.8. After 74 hours, the pH decreased to 6.07 reaching a level of 6.0 by the end of the eleven-day treatment.

The TOC in the sludge exiting the water treatment plant was 44.00 % (± 0.17). During the five-month drying period, the content of organic carbon decreased steadily, reaching 37.3 % (± 1.53) (**Fig. 1**). This decreasing trend has been observed in other studies [11, 20, 21]; however, in the present study the rate was less steep. Furthermore, the TOC levels measured during all the drying period in this study are higher than the observed in other studies [11, 20].

Table 3. Methods used for the physicochemical and microbiological analyses.

| Parameters | Methods |
|---|---|
| Heavy metals (Cd, Hg, , Cu, Ar, Ni, Pb, Se, Zn, Mo) | Atomic absorption spectroscopy |
| Total organic carbon (TOC) | Colorimetry |
| Total phosphorus (P) | Colorimetry |
| Organic nitrogen | Kjeldahl |
| pH | Electrometric 2:5 a 25°C |
| Humidity | Gravimetric |
| Conductivity | Electrometric 1:5 (25°C) |
| Organic matter (MO) | Calculation |
| Total coliforms | Plate-counting in Chromocult agar (Merck) |
| <i>Salmonella</i> sp. | <i>Salmonella</i> in Sewage Sludge (Biosolids) by Modified Semisolid Rappaport-Vassiliadis (MSRV) Medium. |
| Total helminth eggs | Modified Bailenger method |
| Somatic phages | Double-agar-layer method |

The temporal profile for total phosphorus is displayed in **Fig. 2**. Before the second month, total phosphorus remained around 5.21 % (± 0.15). In the second month, the average concentration was reduced to 4.73 % (± 0.05). From the second to the fifth months, the concentration increased steadily until reaching a concentration of 5.08 % (± 0.06). Regarding organic nitrogen, its values fluctuated around 1.91 % (± 0.32). The total phosphorus and organic nitrogen values measured at this plant were higher than those reported by other domestic wastewater treatment plants [11, 21-23]. These high levels are attributable to the main local economic activity of the municipality, which is the production of artisanal dairy products.

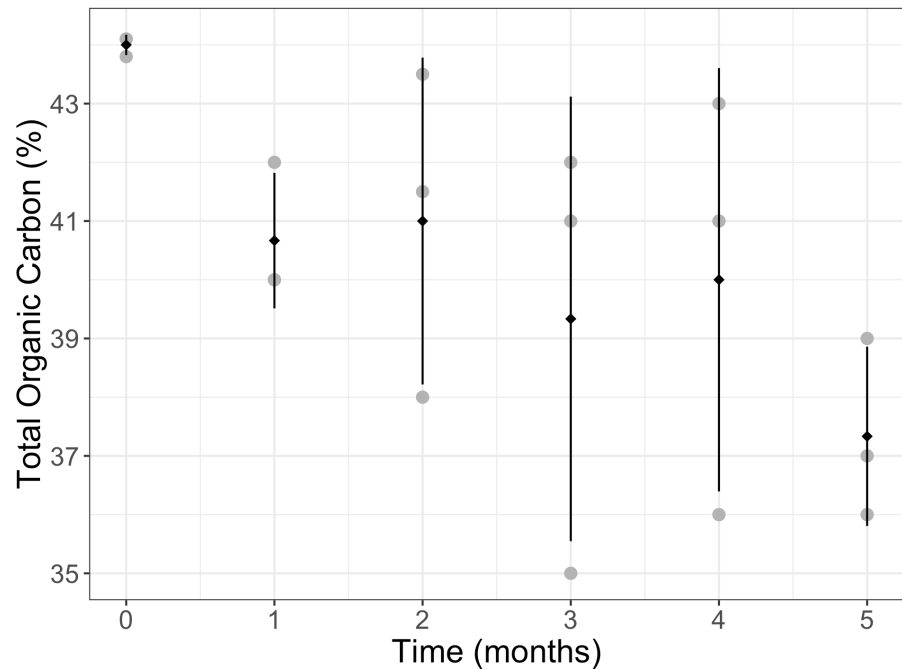


Figure 1. Temporal profile of total organic carbon content. The gray dots represent the observed values, the black points correspond to the average, and the vertical lines indicate the average \pm one standard deviation.

Microbiological analyses showed that the five-month sludge drying process on drying bed had a significant influence on the pathogenic microorganisms. By the fourth month of drying, *Salmonella* sp., was not detected. This absence could be justified by the fact that *Salmonella* sp. is a microorganism sensitive to desiccation and is often outcompeted by other microorganisms [24, 25].

Fecal coliforms decreased from 10.9 log CFU/g (\pm 0.07) to 4.11 log CFU/g (\pm 0.08) during this study. Within the first two months, fecal coliforms decreased more rapidly than in the third to the fifth months (Fig. 3). The pattern of decrease of fecal coliforms concentrations matches that observed for humidity. Humidity was reduced by 88.9% at the end of the second month (Fig. 3); subsequently, it remained relatively constant and reached its lowest point (< 6%) by the end of the fifth month.

Fig. 4 displays the number of total helminth eggs found in the monthly samples. A high prevalence of *Ascaris lumbricoides* (57%) was observed, followed by *Fasciola* sp., *Diphyllobothrium* sp., *Clonorchis* sp., *Schistosoma* spp. and *Hookworm* sp. The mean number of total helminth eggs and its variability decreased with time. An average of 9 (\pm 7.55) total helminth eggs were found in the sludge sample leaving the treatment plant. After five months on the drying bed, only one egg was found on one of the three samples.

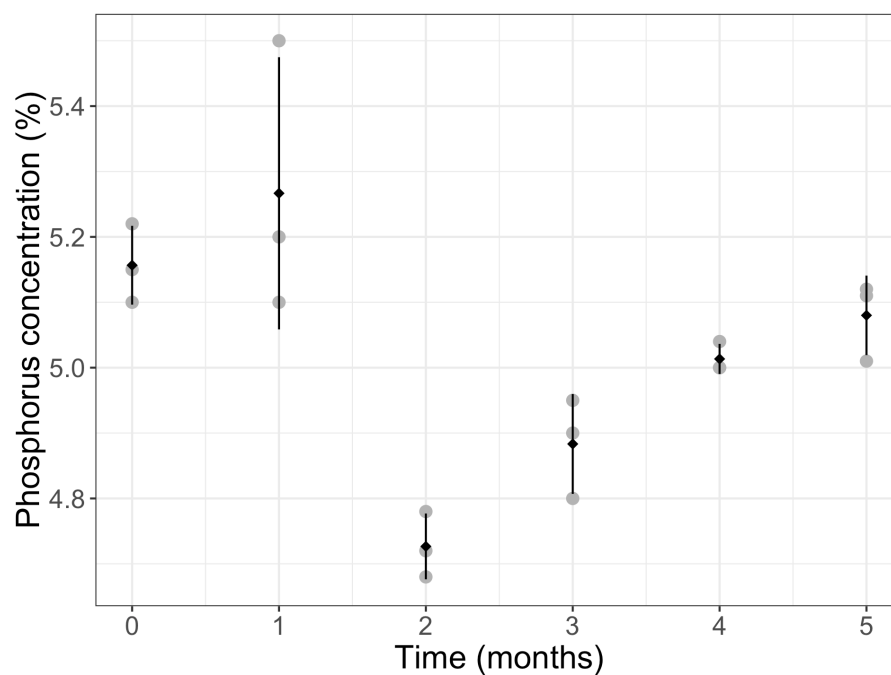


Figure 2. Temporal profile of phosphorus concentration. The black points correspond to the average for each month and the black line represents the average plus and minus the standard deviation.

An average of $4.36 (\pm 0.312)$ log CFU/g somatic phages were observed in the sludge exiting the treatment plant. After two months on drying beds, the log CFU/g somatic phages decreased to $3.52 (\pm 0.03)$. However, after this initial decrease, there was no change in the somatic phages content. These results are consistent with those reported by Shanahan and Campos [24-26].

Physicochemical and microbiological features of sludge under alkaline stabilization

The application of CaO at different doses (7%, 9%, and 13%) produced an increase in the pH of the sludge. The pH of the sludge treated with the highest concentration of CaO (13%) increased to 12 and remained at this level for more than 96 hours (Fig. 5). This treatment complies with one of the alternatives given by the Environmental Protection Agency (EPA) and Colombian regulations to obtain type A sewage sludge [4]. In contrast, the average sludge pHs achieved with CaO concentrations of 9% and 7% were 11.65 and 11.09, respectively.

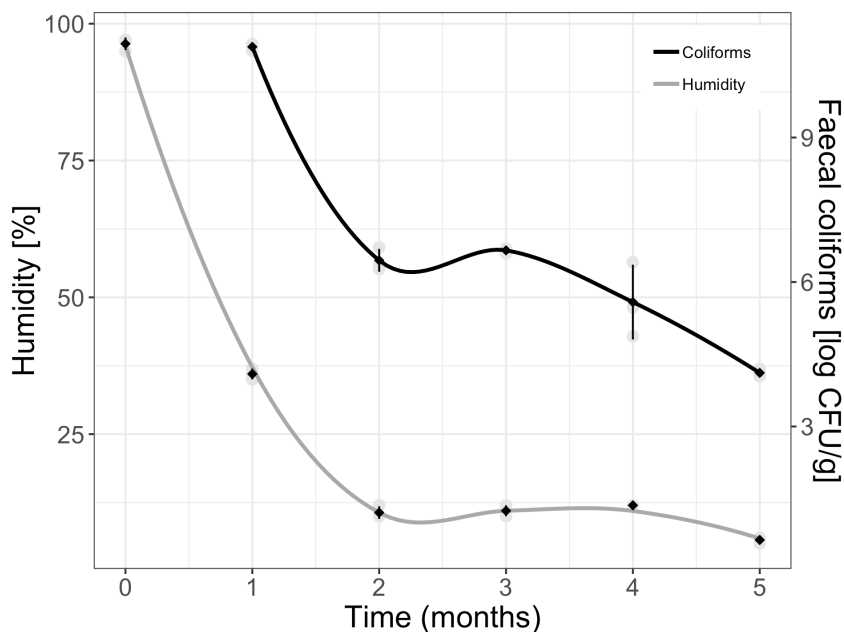


Figure 3. Temporal profile of fecal coliforms and humidity. Black points correspond to the monthly averages and black vertical lines denote the corresponding, plus and minus, standard deviations.

Alcaline treatment with 13 % CaO led to an immediate total reduction of fecal coliforms, *Salmonella* sp., and somatic phages. In the case of helminth eggs, no eggs were detected at time 0 and after 12 hours, but one helminth egg was

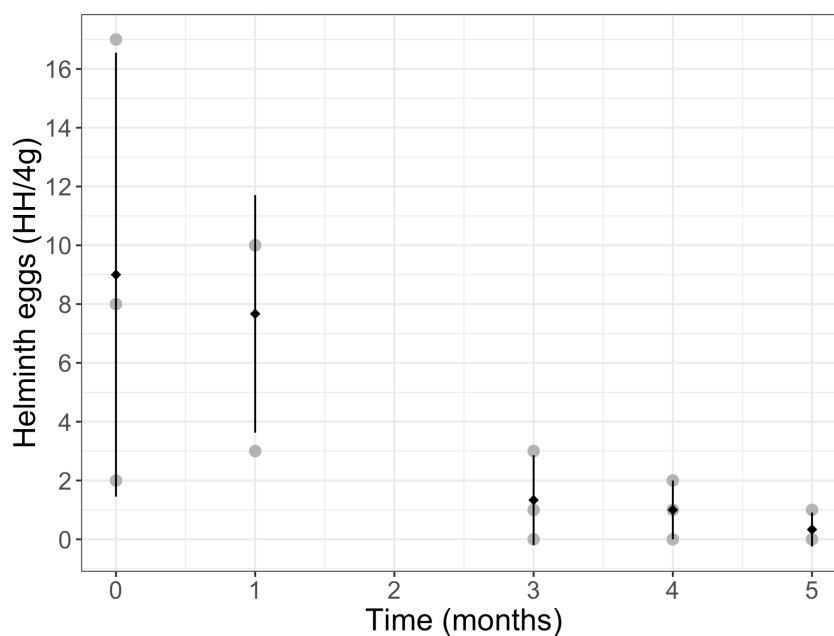


Figure 4. Temporal profile of helminth eggs. Black points correspond to monthly averages and black line denote the corresponding plus and minus standard deviations.

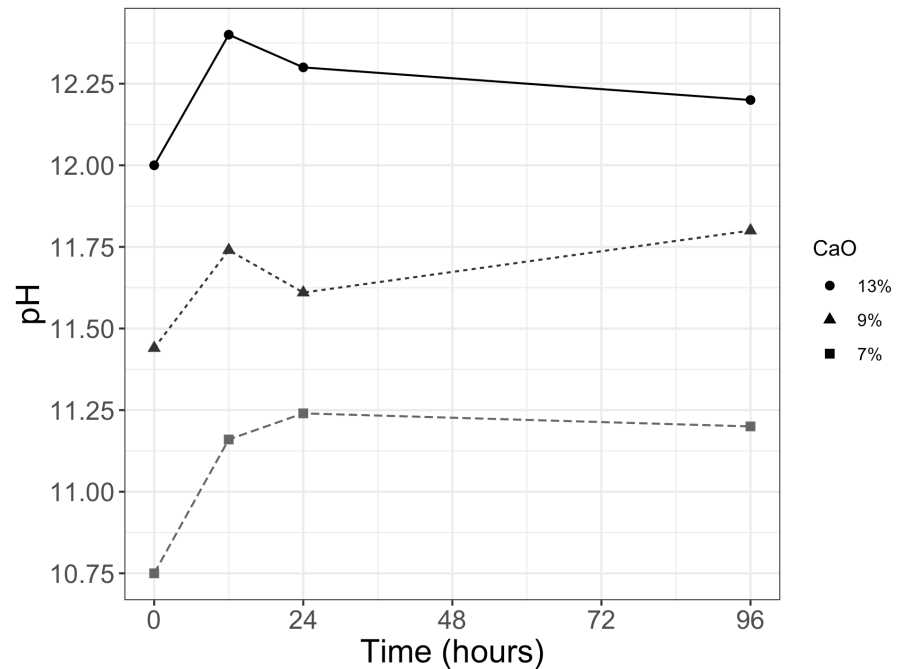


Figure 5. Effect of CaO concentration on sludge pH for 96 hours.

detected in one of the three samples 24 hours after treatment. No helminth eggs were detected 96 hours after treatment. Alkaline treatment with CaO at concentrations of 9 % and 7 % eliminated the fecal contamination indicators after 24 hours and 4 days of exposure, respectively. This elimination is due to the increase in temperature and pH for two or more hours with release of gases and hindering of pathogen reproduction [27]. Both of these concentrations also eliminated *Salmonella* sp., right after application. Helminth eggs were also eliminated with these two CaO concentrations. The application of 9 % CaO resulted in the elimination of all helminth eggs after 12 hours, whereas 7 % CaO eliminated all eggs at 24 hours. Somatic phages were eliminated after 12 hours with both concentrations.

Sludge treatment with 13 % CaO led to results similar to those obtained by Torres *et al* [28] with 15 % CaO, observed helminth egg and fecal coliforms elimination at time zero. These authors also used 8 % CaO and found that helminth eggs were also eliminated at time zero, whereas fecal coliforms were eliminated by the sixth day after application. Silva-Leal *et al* obtained comparable results for helminth eggs with CaO at concentrations of 8 %, 8.5 %, 9.0 %, 9.5 %, and 10 % [29]. Furthermore, contrary to the results reported by Torres *et al* [28] and the present study, Silva-Leal and colleagues did not detect coliforms at the time of CaO application, regardless of its concentration.

Another difference to the present study is that they reached a pH of 12 with all tested CaO concentrations for more than 72 hours as required by EPA.

The treatment with 7 % CaO resulted in sludge phosphorus concentrations of 4.6 % and 4.3 % at the time of application and at 24 h after application, respectively. These concentrations are slightly lower than the ones obtained without alkaline stabilization. On the other hand, in the mixtures treated with 9 % and 13 % CaO phosphorus concentrations at 24 h after application were of 1.0 % and 0.5 %, respectively. The concentrations of organic nitrogen in the sludge mixtures treated with 13 %, 9 %, and 7 % CaO were 1.2 %, 0.9 %, and 1.1 %, respectively. These values are all lower than the obtained without alkaline stabilization.

Based on the physicochemical analysis and microbiological assessment results, we were able to ascertain that sludge dewatering on drying beds is an effective treatment to reduce fecal pollution indicators without altering the sludge's physicochemical properties of agricultural importance. In this study, the biosolid obtained using drying beds was of category B according to Colombian regulations, failing to meet criteria to be considered as a category A biosolid because of reported fecal coliform concentrations. Given that sludge dewatering on drying beds is a process where sunlight and wind stabilize the sludge, it is regarded as a low-cost process with operational simplicity and limited energy consumption. Thus, it is a viable alternative to treat the sludge produced by small WWTPs. However, it requires space and infrastructure that several small plants may not have. On the other hand, alkaline stabilization requires less time and less space; however, the treatment of all the sludge produced will require large quantities of CaO. Therefore, it is desirable to use the minimal concentration of CaO that can reduce fecal pollution indicators. For the small WWTP under study, alkaline stabilization with 7 % CaO can yield A category biosolids.

Conclusions

The physicochemical characterization carried out on the sludge generated in the Sotaquirá WWTP suggests that this sludge has potential agricultural use due to its high content of total organic matter, organic elements (carbon, phosphorus, and nitrogen), and low levels of heavy metals. The two stabilization treatments used allowed to establish a viable alternative in the sanitization of sludge for agricultural applications with some restrictions. In the case of dewatering using drying beds, a category B biosolid was obtained after five months of treatment without affecting the physicochemical characteristics of the sludge. In contrast, alkaline stabilization with CaO concentrations of 7 %, 9 %, and 13 % reduced sludge pathogen loads, successfully yielding biosolids of the A category. Notably, the two low concentrations of CaO did not reach pH levels of 12 units as recommended

by EPA. The time required to achieve a category A biosolid increased when the concentration of CaO was lower. Nonetheless, it is recommended to use 7 % CaO because at this concentration the physicochemical characteristics of the sludge remained unaffected.

Although category A biosolids can be achieved with alkaline stabilization, small WWTPs in Colombia are usually unaware of the potential agricultural use of these biosolids and are not willing to invest in this alternative. Dewatering on drying beds could be the most accessible sludge treatment type for small WWTPs. However, a longer drying time should be considered to reduce the load of pathogenic microorganisms yielding a biosolid suitable for agricultural use.

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Conflict of interest

The authors have no conflicts of interest to declare.

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Evaluación de dos métodos de estabilización de lodos en una planta de tratamiento de aguas residuales en Sotaquirá, Colombia

Resumen: El lodo es un subproducto del tratamiento de aguas residuales y es frecuentemente empleado en procesos agrícolas después de una adecuada estabilización microbiológica. En este estudio se evaluaron dos métodos de estabilización de lodos en la planta de tratamiento de aguas residuales de Sotaquirá, Colombia. Los dos métodos evaluados fueron: deshidratación en lechos de secado y estabilización alcalina. La estabilización a través de deshidratación en lechos de secado se llevó a cabo por cinco meses; la estabilización alcalina se evaluó bajo tres concentraciones de óxido de calcio: 7 %, 9 % y 13 % p/p. Los parámetros de humedad, pH, carbono orgánico, fósforo, nitrógeno orgánico, coliformes fecales, fagos somáticos, huevos de helminto totales y *Salmonella* sp. se monitorearon mensualmente en el lodo en deshidratación, y a las 0, 12, 24 y 96 horas en el lodo bajo estabilización alcalina. Ambos tratamientos eliminaron *Salmonella* sp. del lodo. Los huevos de helminto se eliminaron con estabilización alcalina, y con el tratamiento de deshidratación se redujeron a uno o cero. Los fagos somáticos se eliminaron con estabilización alcalina, pero se redujeron solamente a 3.52 log UFC/g con el método de deshidratación. La deshidratación en lechos de secado produjo biosólidos que pueden utilizarse en restauración de suelos, mientras la estabilización alcalina produjo biosólidos que se pueden usar con propósitos agrícolas. La estabilización alcalina con 9 % y 13 % de óxido de calcio redujo ostensiblemente los contenidos de nitrógeno y fósforo del lodo, mientras que el calcio al 7 % afectó menos la concentración de fósforo. Estos resultados indican que la deshidratación en lechos de secado es un protocolo efectivo de estabilización de lodos, que puede ser implementado en plantas pequeñas de tratamiento de aguas residuales, como la de Sotaquirá, Colombia.

Palabras clave: Lechos de secado, estabilización alcalina; lodo de aguas residuales; biosólidos; patógenos; carbono orgánico.

Avaliação de dois métodos de estabilização de lodos em uma planta de tratamento de águas residuais em Sotaquirá, Colômbia

Resumo: O lodo é um subproduto do tratamento de águas residuais e é frequentemente empregado em processos agrícolas depois de uma adequada estabilização microbiológica. Neste estudo avaliaram-se dois métodos de estabilização de lodos em uma planta de tratamento de águas residuais em Sotaquirá, Colômbia. Os dois métodos avaliados foram: desidratação em leitos de secagem e estabilização alcalina. A estabilização através de desidratação em leitos de secagem foi realizada durante 5 meses; a estabilização alcalina se avaliou em três concentrações de óxido de cálcio: 7 %, 9 %, e 13 % p/p. Os parâmetros de humidade, pH, carbono orgânico, fósforo, nitrogênio orgânico, coliformes fecais, fagos somáticos, ovos de helmintos totais e *Salmonella* sp. Foram monitorados mensalmente no lodo em desidratação, e as 0, 12, 24 y 96 horas nos lodos em estabilização alcalina. Ambos tratamentos eliminaram *Samonella* sp do lodo. Os ovos de helmintos se eliminaram com estabilização alcalina, e com o tratamento de desidratação se reduziram a um ou zero. Os fagos somáticos se eliminaram com estabilização alcalina, mas com o método por desidratação unicamente se reduziram a 3.52 log UFC/g. A desidratação em leitos de secagem produziu biosólidos que podem ser utilizados em restauração de solos, enquanto a estabilização alcalina produziu biosólidos que podem ser usados com propósitos agrícolas. A estabilização alcalina com 9 % y 13 % de óxido de cálcio reduziu ostensivamente os níveis de nitrogênio e fósforo do lodo, enquanto o cálcio a 7 % afetou menos a concentração de fósforo. Estes resultados indicam que a desidratação em leito de secagem é um protocolo efetivo de estabilização de lodos, que pode ser implementado em plantas pequenas de tratamento de águas residuais, como a de Sotaquirá, Colômbia.

Palavras-chave: Leitos de secado; estabilização alcalina; lodo de águas residuais; biosólidos; patógenos; carbono orgânico.

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